

# Greenhouse gas emissions from selected Austrian dairy production systems—model calculations considering the effects of land use change

S. Hörtenhuber<sup>1,2\*</sup>, T. Lindenthal<sup>2</sup>, B. Amon<sup>3</sup>, T. Markut<sup>2</sup>, L. Kirner<sup>4</sup>, and W. Zollitsch<sup>1</sup>

<sup>1</sup>Division of Livestock Sciences, Department of Sustainable Agricultural Systems, University of Natural Resources and Applied Life Sciences Vienna, Gregor-Mendel-Straße 33, A-1180 Vienna, Austria.

<sup>2</sup>Research Institute of Organic Agriculture (FiBL) Austria, Seidengasse 33-35/13, A-1070 Vienna, Austria.

<sup>3</sup>Division of Agricultural Engineering, Department of Sustainable Agricultural Systems, University of Natural Resources and Applied Life Sciences Vienna, Peter Jordan Straße 82, A-1190 Vienna, Austria.

<sup>4</sup>Federal Institute of Agricultural Economics, Marxergasse 2, A-1030 Vienna, Austria.

\*Corresponding author: stefan.hoertenhuber@boku.ac.at

Accepted 4 January 2010

Research Paper

## Abstract

The aim of this study was to analyze various Austrian dairy production systems (PS) concerning their greenhouse gas emissions (GHGE) in a life-cycle chain, including effects of land-use change (LUC). Models of eight PS that differ, on the one hand, in their regional location (alpine, uplands and lowlands) and, on the other hand, in their production method (conventional versus organic, including traditional and recently emerging pasture-based dairy farming) were designed.

In general, the GHGE-reducing effect of a higher milk yield per cow and year in conventional dairy farming cannot compensate for the advantages of organic dairy production which requires lower inputs. This is shown both for GHGE per kg of milk and GHGE per ha and year of farmland. Especially when (imported) concentrates were fed, which had been grown on former forests or grassland, e.g. soybean meal and rapeseed cake, GHGE of conventional dairy farming rose due to the effects of LUC.

GHGE per kg milk varied from 0.90 to 1.17 kg CO<sub>2</sub>-eq for conventional PS, while organic PS on average emitted 11% less greenhouse gases (GHGs), the values ranging from 0.81 to 1.02 CO<sub>2</sub>-eq per kg milk. Within each production method, PS with a higher milk output generally showed better results for GHGE per kg of milk produced than PS with a lower milk output. Nevertheless the latter showed clearly better results for GHGE per ha of land used, ranging from 5.2 to 7.6 Mg CO<sub>2</sub>-eq per ha and year for conventional PS and from 4.2 to 6.2 Mg CO<sub>2</sub>-eq per ha and year for organic PS. The results of this study emphasize the importance of a complete life-cycle assessment in the evaluation of impacts that dairy PS have on the climate.

**Key words:** dairy cow, milk, greenhouse gas emissions, land-use change, mitigation

## Introduction

Agriculture, especially animal husbandry, causes considerable greenhouse gas emissions (GHGE). In the EU-15, agriculture accounted for approximately 10% of total GHGE in 2000<sup>1</sup>. On the one hand, cattle and other ruminants emit relatively large quantities of greenhouse gases (GHGs), particularly methane from enteric fermentation. On the other hand, a large percentage of the agriculturally utilized land in Austria is located in mountainous areas and

uplands<sup>2</sup>. As in other regions that are dominated by grassland, cattle and other ruminants are an essential element of regional agricultural food production. Cessation of (livestock) farming in these regions would therefore have tremendous socio-economic and high ecological costs<sup>2,3</sup>.

Grasslands, pasture and (tropical) forests are vegetations with a high environmental value, with high biodiversity and carbon storage potential<sup>4,5</sup>. Land-use change (LUC), especially in combination with forest clearing in the

tropics, is assumed to cause up to one-quarter of anthropogenic CO<sub>2</sub>-emissions<sup>5</sup>. However, the relative contribution of LUC to GHGE from traditional grassland-based dairy production through imported concentrates has not yet been studied<sup>6</sup>. Another source that is sometimes not considered in estimations of GHGE is the emissions occurring during the rearing period of heifers, on the one hand; on the other hand, beef from calves and cull cows are important by-products from dairy production which result in a relative reduction of GHGE accountable to milk production, but are nevertheless not always considered in respective calculations (e.g. Löthe *et al.*<sup>7</sup>).

Given the regional importance of agriculture in general, and especially of dairy production in alpine regions of Europe, local traditional production systems (PS) must be further developed, including a reduction of their GHGE. Nevertheless, for these PS mitigation options have not yet been extensively analyzed in the literature as most studies have covered more intensive lowland PS which were characterized by greater livestock density and higher quantities of purchased production factors, such as bought-in feed or chemicals (e.g. Thomassen *et al.*<sup>8</sup>, Williams *et al.*<sup>9</sup> and Cederberg and Mattson<sup>10</sup>).

Therefore the goals of this study were to estimate the level of GHGE for selected milk PS in Austria, taking a great number of sources for GHGE into account, to analyze relevant influencing factors and to identify options for their reduction.

## Material and Methods

Models for different PS were built, using MS Excel for calculation and taking into account emissions of methane (CH<sub>4</sub>) and nitrous oxide (N<sub>2</sub>O) from enteric fermentation and from manure management, as well as of CH<sub>4</sub>, N<sub>2</sub>O and carbon dioxide (CO<sub>2</sub>) from soil, from the use of fuels and other energy sources and from production and application of mineral fertilizers and pesticides. Total emissions per cow and year, per kg milk and per ha of farmland used were calculated by adding up the emissions of CH<sub>4</sub>, N<sub>2</sub>O and CO<sub>2</sub> as CO<sub>2</sub>-equivalents (CO<sub>2</sub>-eq). Conversion factors used to calculate the global warming potential are 23 kg CO<sub>2</sub>-eq for 1 kg methane and 296 kg CO<sub>2</sub>-eq for 1 kg N<sub>2</sub>O (100-year-horizon)<sup>11</sup>.

### Modeled dairy PS

Models for Austrian dairy PS were built on farm-level for alpine regions, uplands and lowlands, each one for organically and conventionally managed PS. PS a (alpine) represents a traditional alpine farm at more than 800 m asl and steeply sloping grassland. PS U (uplands) represents a dairy farm in a region of transition from alpine to lowlands with about 20% of arable land and 80% of permanent grassland, while PS L (lowlands) is located in the lowlands, farming on permanent grassland and arable land at a ratio of about 50 : 50. Generally, the intensity of production (i.e.,

among others, stocking rate and milk yield per cow) increases within these PS (PS A < PS U < PS L), while PS UP (uplands, pasture) represents a low-input, pasture-based production system with a relatively high stocking rate. Herein, the appendices 'org' and 'con' are used to further differentiate between organic and conventional PS, respectively. These PS may not necessarily represent the average Austrian dairy farm, but rather represent a wide spectrum of different conditions for dairy production. The great variability in farming conditions in Austrian dairy production are due to the geographic heterogeneity, to the originally small farm size and the different development during the past three decades. According to an analysis, which was based on a comprehensive Austrian agricultural statistical database (Invekos)<sup>12</sup>, about 14% of Austrian dairy farms could be assigned to the PS A<sub>con</sub>, A<sub>org</sub>, UP<sub>con</sub> and UP<sub>org</sub> (between 4500 and 6000 kg milk quota per cow and year in an alpine or upland region), with a relatively high share (26%) of organic farms. Within these 14%, pasture-based PS contribute an estimated 15%. Although currently only relatively few farms exist which are strictly following a pasture-based low-input approach with the highest possible proportion of pasture (up to 60%; Table 1) and seasonal calving, this concept of dairy production is extensively discussed as a future strategy of milk production in permanent grassland regions<sup>13</sup>. PS UP<sub>con</sub> and UP<sub>org</sub> were defined to represent farms following such a strategy. PS U<sub>con</sub> and U<sub>org</sub> represent another 8% of Austrian dairy farms (i.e., with a share of grassland between 65 and 85%), with 16% of farms being organically managed. Furthermore, about 4% of Austrian dairy farms are represented by PS L<sub>con</sub> and L<sub>org</sub> with a milk quota of 6500–8000 kg milk per cow (among them about 9% of organic farms).

Due to differences in management (e.g., number of harvests, dietary proportion and type of bought-in and home-grown feed), milk yield and usage of fuel, mineral fertilizers and pesticides, among other factors, differ between the eight PS. Key characteristics for these PS are shown in Tables 1, 2 and 3.

**Crop and grassland yields, feeding value.** Values for crop yields were derived from Austrian statistical databases<sup>14</sup>, differentiating between conventional and organic production. Yields for grassland were taken from Buchgraber and Gindl<sup>15</sup> and an agricultural national database<sup>16</sup> and were equal for conventional and organic production but were adjusted to the altitude. The main reasons for the assumption of equal forage yields were similar amounts of manure applied, lower gaseous N-losses in organic housing due to higher proportions of pasture and straw-based manure systems (20 and 30% gaseous N-losses from NH<sub>3</sub> and NO<sub>x</sub> emissions for pasture and storage of solid manure, respectively, as compared to 40% N-losses for slurry systems; IPCC<sup>17</sup>, Tables 10.22 and 11.3) and a higher share of legumes in organic grassland and hence higher biological N-fixation (e.g. Rahmann and Böhm<sup>18</sup>). Information on nutrient contents of crops and forage were derived from DLG feed tables<sup>19</sup>, Buchgraber

**Table 1.** Emissions from manure management, direct and indirect soil emissions for the eight PS.

Trait	PS A <sub>con</sub>	PS A <sub>org</sub>	PS UP <sub>con</sub>	PS UP <sub>org</sub>	PS U <sub>con</sub>	PS U <sub>org</sub>	PS L <sub>con</sub>	PS L <sub>org</sub>
Percentage of manure excreted in straw-based systems (%)	43	40	24	24	60	51	60	51
Percentage of manure excreted in slurry-based systems (%)	29	26	16	16	40	33	40	33
Percentage of manure excreted on pasture (%)	28	34	60	60	0	16	0	16
N <sub>2</sub> O—manure (kg cow <sup>-1</sup> year <sup>-1</sup> )	1.2091	1.1280	0.6688	0.6791	1.8388	1.5220	1.9513	1.5688
CH <sub>4</sub> —manure (kg cow <sup>-1</sup> year <sup>-1</sup> )	25.49	23.30	15.67	15.43	37.13	30.32	39.26	31.26
CO <sub>2</sub> -equivalents—manure (kg cow <sup>-1</sup> year <sup>-1</sup> )	944.2	869.8	558.4	555.9	1398.3	1147.9	1480.6	1183.3
N <sub>2</sub> O-emitting soil-N (kg cow <sup>-1</sup> year <sup>-1</sup> )	114.5	110.6	70.1	70.1	112.5	102.4	95.3	98.8
Direct N <sub>2</sub> O soil emissions (kg cow <sup>-1</sup> year <sup>-1</sup> )	2.16	2.16	1.61	1.61	1.93	1.73	1.50	1.67
Indirect N <sub>2</sub> O soil emissions from deposition and leaching (kg cow <sup>-1</sup> year <sup>-1</sup> )	0.62	0.62	0.59	0.59	0.70	0.66	0.77	0.67

et al.<sup>20</sup> and Resch et al.<sup>21</sup>. The management and production factors for the different PS were defined based on Austrian agricultural statistical data<sup>14,16,22,23</sup>, and are described in Tables 2 and 3.

Forage quality depended, among others, on altitude, cutting frequency and time of harvest or grazing, ranging from 5.35 to 6.15 MJ NE<sub>L</sub> ('net energy lactation'<sup>24,25</sup>) per kg of dry matter (DM)<sup>15,21</sup>.

**Feeding regimen and milk yields.** The relative proportion of pasture in the total forage was assumed to decrease from PS UP via PS A to PS U; in PS L<sub>con</sub>, cows were not grazed anymore. The dietary percentage of hay also decreased in the same order, whereas the percentage of silage increased. In general, organic PS fed more hay and used more grazing than conventional PS, as it is shown in Table 2. In the most intensive PS L<sub>con</sub>, about 40% of grass-clover silage were replaced by maize silage. Forage harvest-losses were related to the type of forage fed: pasture 25%, indoor grass feeding 5%, grass silage 20% and hay 30%, on a DM basis<sup>16</sup>.

The annual average percentage of concentrate in the diets was assumed to be between 13% of total feed intake in PS UP (conventional and organic) and 24% in PS U<sub>con</sub> and PS L<sub>con</sub>. In the pasture-based PS UP<sub>con</sub> and PS UP<sub>org</sub>, the concentrate only consisted of grains and mineral premix. Organic concentrate consisted of barley, wheat, faba beans, peas and mineral premix for PS A<sub>org</sub>, PS U<sub>org</sub> and PS L<sub>org</sub>. Conventional concentrate contained barley, wheat, corn, rapeseed cake, extracted soybean meal, faba beans and mineral premix for PS A<sub>con</sub>, PS U<sub>con</sub> and PS L<sub>con</sub>. Forty-five percent of the rapeseed for oil milling (from which rapeseed cake originates as a by-product) was assumed to be imported from abroad (mainly from European countries such as Hungary and Slovakia), the rest was produced in Austria<sup>26</sup>. It was assumed that the production and extraction of soybeans took place mainly in South

America and Germany, respectively<sup>27</sup>. While PS A and PS UP had to buy-in all the concentrates, PS U<sub>con</sub>, PS U<sub>org</sub>, PS L<sub>con</sub> and PS L<sub>org</sub> bought-in 56, 58, 53 and 24% of total concentrates, respectively.

Milk yields ranged from 5500 kg per lactation (organic and conventional PS UP and PS A) to 8000 kg per year (PS L<sub>con</sub>), with an estimated average lifetime performance of 23,650 kg milk, as was recorded for Austria's main breed, Simmental<sup>28</sup>.

Livestock density per hectare (stocking rate) was related to the PS as well as the feeding strategy and was between 1.0 and 1.5 livestock-units of dairy cows per hectare. Internal farmland was assumed to be between 0.67 and 1.0 ha per cow, but due to the demand on land for the production of bought-in feed, total farmland required ranged from 0.84 to 1.23 ha per cow (Table 3). Although cash crops may be produced, particularly in PS U and PS L, only land for cattle feed production was considered herein.

### Sources of emissions

**Enteric fermentation.** Unlike CO<sub>2</sub> emissions from livestock, which are assumed to be zero due to photosynthesis of plants, emissions of CH<sub>4</sub> have to be considered according to IPCC<sup>11</sup>. CH<sub>4</sub> emissions from enteric fermentation were estimated using an equation established by Kirchgeßner et al.<sup>29</sup>:

$$\text{CH}_4 = 63 + 79 \text{ CF} + 10 \text{ NfE} + 26 \text{ CP} - 212 \text{ EE},$$

where 'CH<sub>4</sub>' describes the enteric methane emissions (in g), 'CF' is dietary crude fiber (in kg), 'NfE' is the dietary easily soluble carbohydrates (N-free extracts; in kg), 'CP' is the dietary crude protein (in kg) and 'EE' is the ether extracts (in kg). Table 4 shows some characteristic traits for the rations (Table 2) fed in the eight PS.

**Table 2.** Characteristics of feed production, feed quality and dietary composition for the eight PS.

Characteristic	PS A <sub>con</sub>	PS A <sub>org</sub>	PS UP <sub>con</sub>	PS UP <sub>org</sub>	PS U <sub>con</sub>	PS U <sub>org</sub>	PS L <sub>con</sub>	PS L <sub>org</sub>
Average number of cuts or frequency of pasturage for permanent grassland and clover ley ( <i>n</i> )	2	2	4	4	3	3	4.5	4.5
Gross yields of permanent grassland/clover ley (Mg DM ha <sup>-1</sup> year <sup>-1</sup> )	6.0	6.0	9.0	9.0	8.0/8.5	8.0/8.5	10.0/11.0	10.0/10.5
Average energy density of forage <sup>1</sup> (MJ NE <sub>L</sub> kg <sup>-1</sup> DM)	5.40	5.35	5.90	5.90	5.60	5.60	6.15	5.80
Energy density of concentrate (MJ NE <sub>L</sub> kg <sup>-1</sup> DM)	8.05	8.05	8.00	8.00	8.05	8.05	8.05	8.05
Overall percentage (type <sup>2</sup> ) of forage per kg diet-DM (%)	83 (grass silage, hay, pasture)	81 (hay, pasture, grass silage)	89 (pasture, hay, grass silage)	89 (pasture, hay, grass silage)	76 (grass-clover silage, pasture, hay, fresh grass)	83 (grass-clover silage, pasture, hay, fresh grass)	76 (grass-maize-clover silage)	81 (grass-clover silage, pasture, hay, fresh grass)
Overall percentage of concentrate per kg diet-DM (%)	17	19	11	11	24	17	24	19
Percentage of bought-in feed per cow (% DM)	17	19	11	11	14	10	13	5

<sup>1</sup> Including clover silage and maize silage in uplands and lowlands PS if existing.

<sup>2</sup> Feedstuffs ranked according to their dietary percentage.

**Energy consumption.** The energy directly used on farm for dairy production, as well as the fuels and electric energy that were consumed during the production of mineral fertilizers and pesticides, were considered herein. The greatest share of the electric energy needed on dairy farms is used to cool milk and to produce and supply (concentrate) feed. Therefore, the amount of energy used was related to the annual milk yield, also considering that more feed was required and that higher-mechanized (energy-consuming) housing systems coincided with higher milk yields. A value of 0.05 kWh per kg milk was assumed<sup>30</sup>. All calculations in this model were done per cow and did not account for differences in farm size. Emissions were estimated to be 0.453 kg CO<sub>2</sub>-eq per kWh on average<sup>31</sup>. The amount of fuel used for cultivating the fields was estimated using standard values from a national database<sup>32</sup>, resulting emissions were calculated according to Fehrenbach *et al.*<sup>4</sup>. Additionally, the energy needed in transporting externally produced feedstuffs was also taken into account according to Wilting *et al.*<sup>33</sup>. Mineral fertilizers and pesticides must not be used in the organic PS and were also not used on the grassland of the conventional PS, PS A<sub>con</sub> and PS UP<sub>con</sub>. However, a proportionate input of these factors was accounted for the concentrates imported into the conventional PS.

Emissions occurring during the production of mineral fertilizers and pesticides were derived from Patyk and Reinhardt<sup>34</sup> and Biskupek *et al.*<sup>35</sup>. Table 5 shows energy-related emission factors used herein and the references from which the data were derived. PS A<sub>con</sub> is used as an example to demonstrate how the emissions attributed to these sources were calculated (Table 6).

Construction of machinery and buildings were not included as sources of emissions in the model calculations, as they were expected to be equal for all PS.

**Manure management.** The manure systems were assumed to represent the situation in Austria: 60.7 and 59.7% of the organic and conventional dairy cows, respectively, are housed in straw-based systems<sup>23</sup>. The remaining systems are slurry-based. Therefore, differences occur between the eight PS, according to the proportion of time spent on pasture; according to Amon *et al.*<sup>22,23</sup>, the amount of manure per cow, its organic DM and nitrogen (N) contents, which are in turn related to milk yield and feed intake. Representative data were derived from Gruber and Steinwider<sup>36</sup>, the amounts of manure were calculated to be between 19.4 (for both PS A and PS UP) and 22.4 (for PS L<sub>con</sub>) Mg per cow and year. The content of volatile solids excreted daily (DM) was calculated to be between 4.0 and 4.6 kg per cow, based on Schechter<sup>37</sup>. The quantity of nitrogen excreted was estimated to vary from 86.2 to 100.6 kg per cow and year if a moderate N-content of feed is assumed<sup>38</sup>. Based on these values, the amounts of CH<sub>4</sub> and N<sub>2</sub>O emitted were estimated according to IPCC<sup>17</sup> (tier 2; equations 10.23 and 10.25) and are given in Table 1, using methane conversion factors of 0.3, 0.04 and 0.015 for slurry, farmyard

**Table 3.** Key characteristics of the eight PS.

Characteristic	PS A <sub>con</sub>	PS A <sub>org</sub>	PS UP <sub>con</sub>	PS UP <sub>org</sub>	PS U <sub>con</sub>	PS U <sub>org</sub>	PS L <sub>con</sub>	PS L <sub>org</sub>
Regional location, production method	Alpine, conventional	Alpine, organic	Uplands pasture-based, conventional	Uplands pasture-based, organic	Uplands, conventional	Uplands, organic	Lowlands, conventional	Lowlands, organic
Stocking density (cow-LU ha <sup>-1</sup> )	1.0	1.0	1.4	1.4	1.2	1.1	1.5	1.2
Internal farmland required per cow <sup>1</sup> (ha)	1.0	1.0	0.71	0.71	0.83	0.91	0.67	0.83
Total farmland required per cow <sup>2</sup> (ha)	1.23	1.34	0.84	0.88	1.11	1.10	0.94	0.92
Permanent grassland, proportion of agricultural land (%)	100	100	100	100	80	80	50	56
Arable land, proportion of agricultural land (%)	0	0	0	0	20	20	50	44
Crop rotation on arable land and percentage of the crops (%)	–	–	–	–	Clover ley (20), wheat (25), barley (55)	Clover ley (25), barley (34), faba beans (25), wheat (16)	Clover ley (20), maize (35), barley (30), wheat (15)	Clover ley (25), barley (34), faba beans (25), wheat (16)
Annual milk yield per cow (kg)	5500	5500	5500	5500	7000	6500	8000	7000
Milk yield per ha of total farmland (kg)	4475	4103	6576	6223	6304	5913	8542	7606
Total farmland required per 1000kg milk (ha)	0.223	0.244	0.152	0.161	0.159	0.169	0.117	0.131

<sup>1</sup> For the production of homegrown feedstuffs.<sup>2</sup> For the production of homegrown plus bought-in feedstuffs.

**Table 4.** Nutrient intake and their relative dietary proportions for the eight PS.

Nutrient	PS A <sub>con</sub>	PS A <sub>org</sub>	PS UP <sub>con</sub>	PS UP <sub>org</sub>	PS U <sub>con</sub>	PS U <sub>org</sub>	PS L <sub>con</sub>	PS L <sub>org</sub>
CF (kg per day; [%])	3.43 [24]	3.47 [24]	3.44 [25]	3.44 [25]	3.27 [22]	3.51 [23]	2.98 [18]	3.21 [22]
NfE (kg per day; [%])	7.09 [50]	7.25 [51]	6.66 [48]	6.66 [48]	7.61 [50]	7.47 [49]	9.08 [56]	7.12 [50]
CP (kg per day; [%])	1.98 [14]	1.91 [13]	2.03 [15]	2.03 [15]	2.34 [15]	2.26 [15]	2.31 [14]	2.18 [15]
EE (kg per day; [%])	0.41 [3]	0.36 [3]	0.34 [2]	0.34 [2]	0.48 [3]	0.44 [3]	0.52 [3]	0.42 [3]
Ash (kg per day; [%])	1.34 [9]	1.29 [9]	1.41 [10]	1.41 [10]	1.45 [10]	1.52 [10]	1.25 [8]	1.41 [10]

CF, dietary crude fiber; NfE, dietary easily soluble carbohydrates; CP, dietary crude protein; EE, ether extracts.

**Table 5.** Overview on used emission factors.

Process	Unit	Emission factor	Reference
Fuels—emissions from supply chain and consumption	kg CO <sub>2</sub> -eq l <sup>-1</sup> diesel	3.2066	Fehrenbach <i>et al.</i> <sup>4</sup>
Electric energy—emissions from supply chain	kg CO <sub>2</sub> -eq kWh <sup>-1</sup>	0.453	Ecoinvent <sup>31</sup>
Production mineral fertilizer—nitrogen (N)	kg CO <sub>2</sub> -eq kg <sup>-1</sup> N-mineral	7.470	Patyk and Reinhardt <sup>34</sup>
Production mineral fertilizer—phosphorus (P <sub>2</sub> O <sub>5</sub> )	kg CO <sub>2</sub> -eq kg <sup>-1</sup> P <sub>2</sub> O <sub>5</sub> -mineral	1.176	Patyk and Reinhardt <sup>34</sup>
Production mineral fertilizer—potassium (K <sub>2</sub> O)	kg CO <sub>2</sub> -eq kg <sup>-1</sup> K <sub>2</sub> O-mineral	0.664	Patyk and Reinhardt <sup>34</sup>
Production pesticides	kg CO <sub>2</sub> -eq kg <sup>-1</sup> pesticide	5.369	Biskupek <i>et al.</i> <sup>35</sup>

manure and pasture, respectively. Emission factors used for calculation of N<sub>2</sub>O were 0.02 and 0.001 for farmyard manure and slurry, respectively<sup>17</sup>.

**Soil N<sub>2</sub>O.** Direct N<sub>2</sub>O emissions were calculated based on the amount of nitrogen introduced into the soil (IPCC-tier 2, equation 11.2)<sup>17</sup>. Amounts of N from mineral fertilizers, mineralization, manure and crop residues were multiplied by the default emission factors of 0.01 kg N<sub>2</sub>O-N per kg of N applied; a factor of 0.02 kg N<sub>2</sub>O-N was used for each kg of N excreted by cows on pasture. Indirect soil emissions from deposited nitrogen and leaching were estimated according to IPCC<sup>17</sup> (default values in Tables 10.22 and 11.3). In Table 1, amounts of N and related N<sub>2</sub>O-emissions are shown for the different PS. N<sub>2</sub>O emitted during cultivation of bought-in concentrates was included, because of its relevance for the N-balance of the total dairy supply chain. Due to a low demand on farmland and less grazing, soil N<sub>2</sub>O-emissions per cow seem to be lower for more intensive PS. On the contrary, emissions from manure are higher for more intensive systems with little or no grazing (Table 1).

**LUC and changes in soil organic carbon stocks.** LUC for soybean production was calculated according to statistics on imports (98% of imported soybean meal originated from South America<sup>27</sup>) and based on estimates for the conversion of savannah-type vegetation into soybean fields<sup>4</sup>. The emissions from LUC, which may be even higher in the case of rainforest clearance, were calculated depending on loss of carbon from the soil and aboveground biomass and were allocated to soybean oil and extracted soybean meal based on their caloric values.

As supported by data from the European Environment Agency (EEA)<sup>39</sup>, LUC from grassland to arable land could be neglected for Austria, but had to be considered for imports of rapeseed from Eastern and Central Europe. An

equivalent of 53% of LUC-related GHGE was estimated for the CO<sub>2</sub>-eq value of rapeseed cake by Fehrenbach *et al.*<sup>4</sup>, which was used as feedstuff for dairy cattle, because 45% of the rapeseed further processed in Austria was imported mainly from Eastern and Central Europe in 2004<sup>26</sup>. Consequently, LUC-related GHGE of 5.41 and 0.40 kg CO<sub>2</sub>-eq were calculated for soybean meal and rapeseed cake, respectively, in addition to emissions from cultivating, transport and processing.

CO<sub>2</sub> sequestered into soil or released from soil organic carbon stocks was calculated according to Küstermann *et al.*<sup>40</sup> for Bavaria, Germany, where on-site conditions can be expected to be very similar to those in Austria. CO<sub>2</sub>-emissions and -sequestration occur for conventionally (+202 kg ha<sup>-1</sup> a<sup>-1</sup>) and organically managed soils (−400 kg ha<sup>-1</sup> a<sup>-1</sup>), respectively, due to differences in crop rotation and manure management. An even higher sequestration rate of 575 kg CO<sub>2</sub> ha<sup>-1</sup> a<sup>-1</sup> was observed in a previous long-term study in Switzerland<sup>41</sup>, where on-site conditions could also be expected to be similar to those in Austria.

It was estimated that 1 kg of conventional concentrate caused emissions of 0.05 kg CO<sub>2</sub>-eq from soil organic carbon changes, based on +202 kg CO<sub>2</sub>-eq ha<sup>-1</sup> a<sup>-1</sup> and an average grain yield of 4000 kg ha<sup>-1</sup> a<sup>-1</sup>. One kilogram of an organic concentrate was expected to be related to a sequestration of 0.111 kg CO<sub>2</sub>-eq (−400 kg CO<sub>2</sub>-eq ha<sup>-1</sup> a<sup>-1</sup> and an average grain yield of 3600 kg ha<sup>-1</sup> a<sup>-1</sup>). Due to its long history of relatively constant management, it was assumed that Austrian alpine grassland is at an equilibrium state and that its soils did not emit or sequester further CO<sub>2</sub><sup>42</sup>.

**Rearing phase and beef as a by-product.** The rearing phase of dairy cows prior to first calving has to be considered as an important source of GHGE, together with

Table 6. Consumption of energy and inputs per cow and year: PS A<sub>con</sub> as an example.

Sources	Demand for additional		Demand for electric energy (kWh)	Emissions from fuels (kg CO <sub>2</sub> -eq)	Emissions from mineral fertilizer (kg CO <sub>2</sub> -eq)	Emissions from pesticides (kg CO <sub>2</sub> -eq)	Emissions from electric energy (kg CO <sub>2</sub> -eq)
	mineral fertilizer <sup>1</sup> (kg N/P <sub>2</sub> O <sub>5</sub> /K <sub>2</sub> O)	pesticides (kg)					
Reference	ÖKL <sup>32</sup> , Wilting et al. <sup>33</sup>	BMLFUW <sup>16</sup>	Öhlinger et al. <sup>30</sup>	Fehrenbach et al. <sup>4</sup>	Patyk and Reinhardt <sup>34</sup>	Biskupek et al. <sup>35</sup>	Ecoinvent <sup>31</sup>
Forage (homegrown)	58.2	9.8/0/85.0	—	186.6	129.6	0	—
Concentrate (bought-in)	17.9	18.6/7.1/6.1	—	57.4	149.6	4.8	—
Transports	13.3	—	—	42.6	—	—	—
Housing	—	—	275	—	—	—	124.6

<sup>1</sup> Taking into account the amount of nutrients in the manure.

beef as the by-product from dairy production having a related mitigating effect for GHGE per kg of produced milk.

In the model calculations, the GHGE during the rearing phase were calculated for each PS based on the average emissions per MJ NE<sub>L</sub> consumed during the rearing phase. In the same way, emissions for a growing–fattening heifer were calculated as a standard of comparison for each PS, based on the emissions per MJ NE<sub>L</sub>. The mitigating effect of beef as a by-product from cull cows and newborn calves (50% bodyweight estimated as carcass) was calculated, using growing–fattening heifer as a model for beef from each PS. By considering all sources of GHGE except electric energy for cooling milk, 1 MJ NE<sub>L</sub> was burdened with 0.15 kg CO<sub>2</sub>-eq (PS L<sub>org</sub>) to 0.19 kg (PS A<sub>con</sub>). The total requirement for energy during a dairy cows' rearing phase was assumed to be about 31,000 MJ NE<sub>L</sub> with an age at first calving of 28 months<sup>28,43</sup>. As a consequence, emissions during the rearing phase were calculated to vary from 4.7 for PS L<sub>org</sub> to 5.9 Mg CO<sub>2</sub>-eq per cow in PS A<sub>con</sub>. Emissions during the growing–fattening of a heifer (about 20,000 MJ NE<sub>L</sub> required for 600 kg final body weight) were estimated to vary from 3.0 to 3.8 Mg CO<sub>2</sub>-eq per head; the mitigating effect of beef from slaughter cows and newborn calves was therefore calculated to be between 3.1 and 3.9 Mg CO<sub>2</sub>-eq per cow.

## Results and Discussion

### GHGE from dairy PS

#### Emissions related to input factors and milk yield.

Total emissions per cow (related to a lifetime milk yield of 23,650 kg) are presented in Table 7. The majority of GHGE evolved from enteric fermentation (40–62%), while the use of fuels and energy (in total 5–9%) as well as production of external inputs such as mineral fertilizers and pesticides (up to 7%) contributed relatively little. In Figure 1, GHGE from soil (N<sub>2</sub>O), from fuels, fertilizers and pesticides used, were aggregated to GHGE accountable to forage and concentrate supply. As the same sources—except electric energy for cooling milk—contribute to emissions during the rearing period and to the mitigation of emissions indirectly caused by the by-product beef, total GHGE from enteric fermentation are actually higher than presented in Figure 1 and Table 7. For example, total GHGE from enteric fermentation are 53% of total emissions (i.e., 3326 kg CO<sub>2</sub>-eq per year) as compared to 49% (i.e., 3100 kg CO<sub>2</sub>-eq annually) which are attributed to one productive year of a dairy cow.

In previous calculations, which focused on the relation between GHGE and milk yield, emissions which originated from the rearing phase were not always taken into account (e.g., Löthe et al.<sup>7</sup>). Although milk yield remains to be an essential factor, the rearing phase is equally important, as a heifer needs nearly the same amount of energy for growth and maintenance as a cow during one lactation. Additionally, the production of beef as a by-product of dairy

**Table 7.** Emissions of GHGs (kg CO<sub>2</sub> eq cow<sup>-1</sup> year<sup>-1</sup>; [%]) from different sources for the eight PS, including the source 'rearing phase' and beef as a by-product.

Source of emission	PS A <sub>con</sub>	PS A <sub>org</sub>	PS UP <sub>con</sub>	PS UP <sub>org</sub>	PS U <sub>con</sub>	PS U <sub>org</sub>	PS L <sub>con</sub>	PS L <sub>org</sub>
Rearing phase—proportional emissions per lactation	1359 [21]	1167 [21]	1210 [21]	1198 [23]	1686 [23]	1314 [22]	1679 [23]	1359 [24]
Enteric fermentation	3100 [48]	3215 [57]	3209 [57]	3209 [62]	2995 [42]	3205 [54]	2841 [40]	2988 [52]
Manure	944 [15]	870 [16]	558 [10]	556 [11]	1398 [19]	1148 [19]	1481 [21]	1183 [21]
Forage—fuels	187 [3]	179 [3]	133 [2]	133 [3]	189 [3]	199 [3]	165 [2]	193 [3]
Forage—inputs	130 [2]	0 [0]	264 [5]	0 [0]	167 [2]	0 [0]	140 [2]	0 [0]
Forage—direct N <sub>2</sub> O	547 [8]	561 [10]	564 [10]	564 [11]	409 [6]	426 [7]	223 [3]	281 [5]
Concentrates—fuels	100 [2]	106 [2]	37 [1]	58 [1]	161 [2]	125 [2]	226 [3]	177 [3]
Concentrates—inputs	154 [2]	0 [0]	104 [2]	0 [0]	125 [2]	0 [0]	166 [2]	0 [0]
Concentrates—direct N <sub>2</sub> O	86 [1]	78 [1]	58 [1]	58 [1]	150 [2]	122 [2]	257 [4]	182 [3]
Indirect soil emissions—N <sub>2</sub> O	183 [3]	179 [3]	176 [3]	176 [3]	206 [3]	196 [3]	228 [3]	199 [3]
Changes in soil organic carbon	48 [1]	-113 [-2]	37 [1]	-81 [-2]	73 [1]	-109 [-2]	77 [1]	-121 [-2]
Production of energy in housing	125 [2]	125 [2]	125 [2]	125 [2]	159 [2]	147 [2]	181 [3]	159 [3]
By-product beef—proportional emissions per lactation	-895 [-14]	-772 [-14]	-801 [-14]	-793 [-15]	-1116 [-16]	-870 [-15]	-1112 [-15]	-899 [-16]
LUC	389 [6]	0 [0]	0 [0]	0 [0]	589 [8]	0 [0]	628 [9]	0 [0]
Total GHGE	6452 [100]	5594 [100]	5675 [100]	5203 [100]	7190 [100]	5902 [100]	7181 [100]	5699 [100]
Total GHGE related to milk yield (kg CO <sub>2</sub> eq kg <sup>-1</sup> milk)	1.173	1.017	1.032	0.946	1.027	0.908	0.898	0.814
Total GHGE related to total farmland (kg CO <sub>2</sub> eq ha <sup>-1</sup> )	5246	4175	6756	5913	6477	5365	7639	6195

production, lifetime performance and the number of lactations that a cow lasts, are important factors as well. Following Fürst<sup>44</sup>, conventional and organic Austrian dairy cows are expected to produce about the same amount of milk in their lifetime, but with the latter being different.

Generally, PS with a higher output of milk showed higher GHGE per cow and year but were superior over PS with a lower output if emissions were expressed per kg of milk. GHGE per kg of conventional and organic milk were between 0.90 and 1.17 kg CO<sub>2</sub>-eq per kg of milk and between 0.81 and 1.02 kg CO<sub>2</sub>-eq per kg of milk, respectively (Table 7, Fig. 1). These numbers illustrate that differences may be smaller between production methods (i.e. conventional versus organic) than between regions: on average, organic PS showed about 11% lower GHGE per kg milk than comparable conventional PS, while the relative difference between the lowest emissions from the PS L and the highest emissions from the PS A amounted to 22%. Generally, the higher the dietary energy and nutrient density, the higher the milk yield, which results in lower GHGE per kg of milk due to reduced enteric fermentation and performance-related degeneration.

As the nutritional value of forages and concentrates is routinely characterized by proximate analysis in Austria, enteric fermentation was calculated according to the regression equation published by Kirchgeßner *et al.*<sup>29</sup>. In order to check these estimates, a regression equation derived by Hindrichsen *et al.*<sup>45</sup> was used. It was found that estimates resulting from the equation by Kirchgeßner *et al.* were on average 8% higher than those resulting from the equation of Hindrichsen *et al.*, with the difference being greater for rations that are rich in crude fiber and being lower for rations rich in soluble carbohydrates.

**The role of LUC.** Another highly relevant source of GHGE has to be taken into account whenever conventional PS utilize soybean meal originating from South America: the production of soybeans is linked to an LUC from former savannah-type vegetation into arable land. In contrast to most previous calculations of GHGE and life-cycle assessments which did not incorporate LUC in their calculations (e.g., Lehuger *et al.*<sup>46</sup>), LUC was taken into account in this study (Table 7, Figs. 1 and 2) due to its high relevance on the global scale<sup>4,5,11</sup>. Estimates for GHGE from LUC vary considerably, e.g., about 5 kg CO<sub>2</sub>-eq per kg of Argentinean soybean meal<sup>47</sup> as compared to more than 10 kg CO<sub>2</sub>-eq per kg of Brazilian soybean meal. The latter occurs if soybean is cultivated on deforested land (calculated from Renewable Fuels Agency<sup>48</sup>). Herein, a change from extensive grassland (savannah) to arable land was assumed<sup>4</sup>, but without deforestation. Since 2000, the area for soybean cultivation expanded predominantly on deforested land<sup>49</sup> and therefore emissions from LUC may actually be even higher than calculated by both Fehrenbach *et al.*<sup>4</sup> and within this study. Nevertheless, even with relatively low estimates for GHGE from LUC, emissions per kg milk will be lower for PS which do not import soybean meal into their



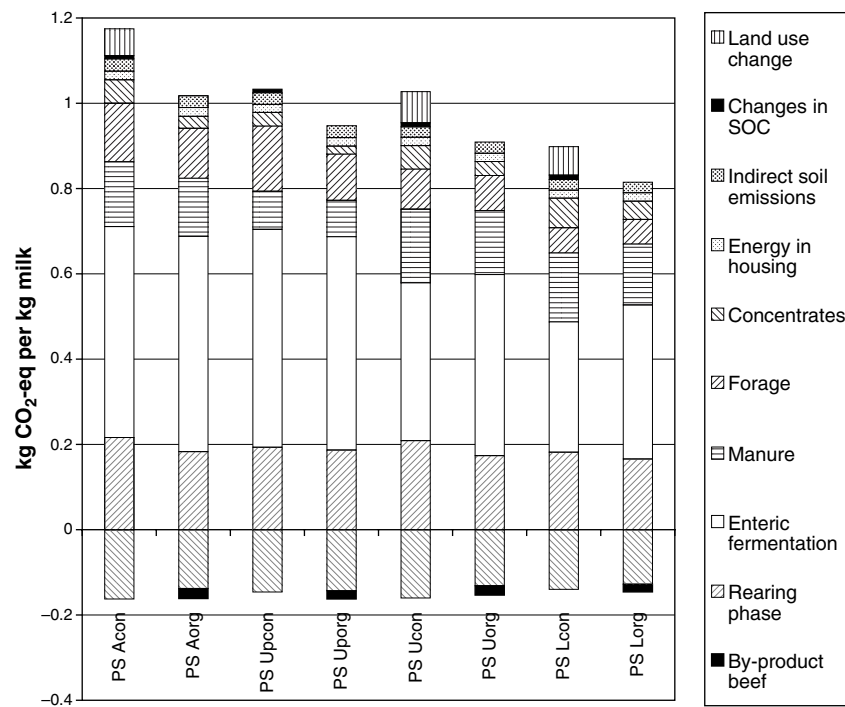


Figure 1. GHGE (kg CO<sub>2</sub>-eq) per kg milk for the eight PS.

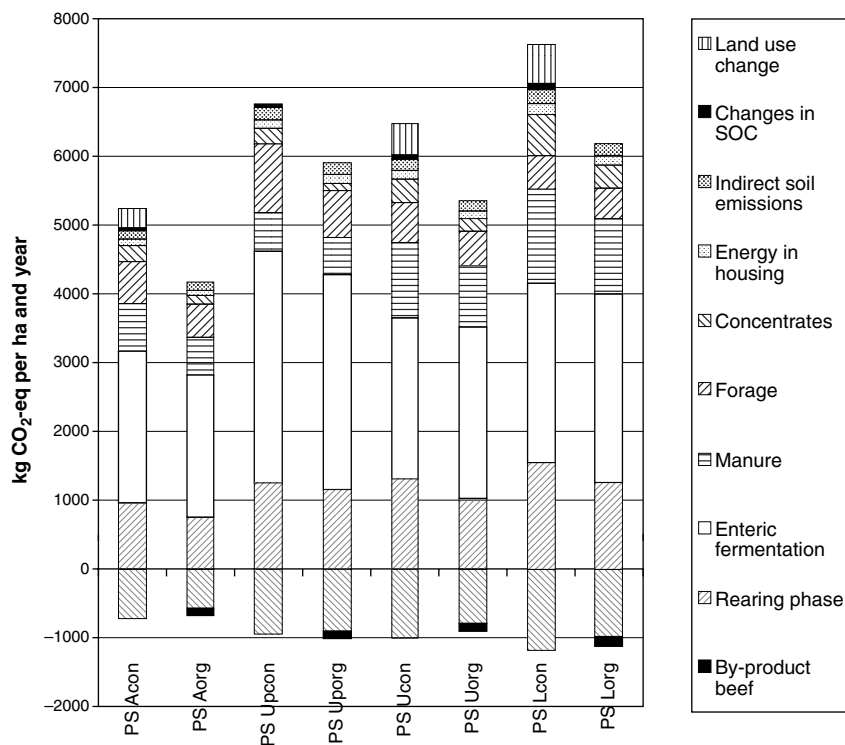


Figure 2. GHGE (Mg CO<sub>2</sub>-eq) per ha and year of total farmland for the eight PS, including the source, rearing phase, and beef as a by-product.

system, but utilize homegrown or locally produced protein sources such as grain legumes or oilseed cakes. This contributes to the average difference in GHGE of 11% between conventional and organic PS studied herein.

About 8% of total GHGE in conventional farming (average over all conventional PS excluding PS UP<sub>con</sub>) result from LUC, mainly (93%) for extracted soybean meal originating from South America.

### Emissions and regional location of dairy production.

The difference in GHGE between alpine and lowlands PS with both conventional and organic management is mainly due to higher milk yields in the lowlands as well as a decreased enteric fermentation as a consequence of increased dietary energy density. Pasture-based PS (UP) emit relatively low amounts of GHGE per kg milk, despite their comparably low milk yield of 5500 kg per cow and year. This is caused by several factors, including reduced emissions from manure in housing (with cows being on pasture for 60% of the time) and the use of low amounts of concentrates. On the contrary, GHGE from enteric fermentation are slightly higher as compared to PS that use less hay but more silage. Pasture-based systems also have higher GHGE from forage because of a doubled rate of N<sub>2</sub>O from N excreted during grazing<sup>17</sup>, as compared to N spread as manure. Both the conventional and organic PS UP show a high productivity per ha of farmland required and low GHGE, which are similar to those of PS U.

Only very few studies are available about GHGE from dairy PS located in alpine or nearby regions (e.g. Weiske *et al.*<sup>50</sup> and Olesen *et al.*<sup>51</sup>). In contrast to the results presented herein, Weiske *et al.*, using a model based on Olesen *et al.*, reported generally higher GHGE per kg milk of between 1.2 and 2.0 kg CO<sub>2</sub>-eq although LUC was not included in their calculations. The differences may be due to the lower milk yield and a lack of differentiation between organic and conventional PS in milk yield as assumed by Weiske *et al.*, a greater number of input factors considered (e.g., the production of mineral premix and seeds), but also disregarding the emission-mitigating effects of by-products.

Great differences occur between PS when GHGE are related to the area of farmland used (Fig. 2). From this perspective, low-input systems show clearly better results. Organic PS need more area per cow due to lower yields especially from arable land, and at the same time show less GHGE per cow. As presented in Figure 2, GHGE per hectare of total farmland vary from 5.2 to 7.6 Mg CO<sub>2</sub>-eq and from 4.2 to 6.2 Mg CO<sub>2</sub>-eq for conventional and organic PS, respectively. Olesen *et al.*<sup>51</sup> reported comparable results of 8.7 Mg CO<sub>2</sub>-eq for conventional PS and 6.0 Mg CO<sub>2</sub>-eq for organic PS calculated according to an IPCC tier 2 methodology<sup>52</sup>.

### Mitigation options for dairy PS

A number of possibilities exist to mitigate GHGE, some concerning the animals, e.g., their genetics for milk yield, but most address management practices. Because of the complex interactions between the various elements of dairy PS and the factors influencing them, any measures intended to reduce the GHGE must be thoroughly examined, as they may exert effects that eventually counteract the intended mitigation.

**Feed quality and LUC.** One of the most effective strategies to reduce emissions is to increase the energy density of the diet, which usually results in the suggestion to increase the proportion of dietary concentrates. On the one hand, this will lead to a significant decrease of GHGE from enteric fermentation (e.g., -6.5% for PS U<sub>con</sub> as compared to PS U<sub>org</sub> if equal milk yields and energy intakes are assumed). On the other hand, increased GHGE from soils and from the use of fertilizers coincide with this mitigating effect and partially counteract it. If an increased use of concentrates is accompanied by LUC (i.e., by converting grassland or pasture into arable land), the mitigating effect of using concentrates turns into an aggravating effect, as shown herein. Another limitation of this practice is the growing probability for digestive and other health disorders associated with increased levels of concentrates in the diet of dairy cows<sup>53,54</sup>. Due to the core role of forages in grassland-based PS (herein, diets were assumed to consist of 76–87% of forage), improving the nutrient density of forage should be prioritized in order to reduce GHGE from enteric fermentation. The greatest effect may be achieved by reducing crude fiber in forage by earlier harvesting (or grazing), but this option is also limited for reasons of grassland ecology<sup>55,56</sup>. According to the calculations conducted within this study, an increase of 0.1 MJ NE<sub>L</sub> per kg forage DM will lead to a reduction of total GHGE of about 1.5%.

**Lifetime performance.** Reducing age at first calving and thereby the rearing phase as well as decreasing the number of lactations in which a constant lifetime performance is yielded would result in a reduced demand for dietary energy and therefore in a reduction of GHGE. On the contrary, a decreased age at first calving and a continuing increase in milk yield per lactation may negatively affect lifetime performance, the number of lactations and the number of offspring per cow<sup>57</sup>. However, an improved lifetime performance (together with a constant or even increasing number of lactations) may be an effective way to decrease GHGE, because emissions from the rearing period will be distributed over a greater amount of milk: GHGE per kg of milk would be reduced by 1.4% on average if lifetime performance was increased by 5000 kg (i.e., from a current value of 23,650 to 28,650 kg for Austria).

**Manure management.** Furthermore, great potential for reduction of GHGE arises from changes in manure management. The PS in this study were assumed to represent the Austrian situation, where 59.7 and 60.7% of the dairy cows were kept in straw-based systems on conventional and organic farms, respectively<sup>23</sup>. All other cows were kept in systems with slurry production. GHGE from cows' excreta were lower for housing systems with straw litter than for slurry-based systems. A change toward straw-based systems would therefore be desirable because of reduced GHGE and also of a potentially beneficial effect on animal welfare<sup>58</sup>. GHGE per kg milk may be mitigated by 0.9% if a further 10% of dairy barns are

changed from slurry- to straw-based systems. However, this is not reflected by the current trend in Austria, where most of the newly built housing systems are slurry based<sup>22</sup>. Moreover, the separation of slurry into solid and liquid phases, as well as an aeration of slurry and composting of farmyard manure would contribute to a mitigation of GHGE, but may lead to increased ammonia emissions<sup>59</sup>. Over all PS covered herein, slurry separation and slurry aeration would reduce total GHGE by 1.8% on average.

When calculating GHGE from manure falling on pasture, only CH<sub>4</sub>, but not N<sub>2</sub>O, was taken into account, as the latter was already included in the emissions from soil<sup>17</sup>. Overall, pasture-based systems can be considered not only as animal friendly but also as favorable from the point of view of GHGE, as they are emitting less GHG than any other housing systems. On average, cows that spend 10% of their annual time budget on pasture emit 2.4% less GHG as compared to cows that are confined all year round.

Another very substantial reduction potential arises from anaerobic fermentation of manure for biogas production<sup>60</sup>, whereby GHGE per kg milk can be decreased by about 5%. Additionally to this direct reduction, the substitution of fossil fuels by biogas could contribute to a further decrease, as 1 Mg of cattle manure can be transformed into 32 kWh of electric energy in a biogas plant<sup>61</sup>. For the PS covered in this study, this would lead to a reduction of total GHGE per kg milk by 7%. Depending on the degree of utilization of the heat that emerges in a biogas plant and the potential substitution of fossil energy for heating, a further potential reduction of GHGE arises.

**Utilization of oil seeds for feed and biofuels.** The utilization of by-products from the production of biofuels in livestock nutrition is frequently advocated as a contribution to improved sustainability of agricultural production (e.g., UN-Energy<sup>62</sup>). Nevertheless, certain energy crops, such as rapeseed, are frequently reported to cause higher emissions of N<sub>2</sub>O than assumed by IPCC and as used in this study and in most previous calculations and reports (e.g., The Royal Society<sup>63</sup>)<sup>64</sup>. According to Crutzen et al.<sup>64</sup>, the N<sub>2</sub>O-emissions from rapeseed are 3–5 times higher than reflected by IPCC<sup>17</sup> default values and current state of life cycle analysis. Assuming five times higher N<sub>2</sub>O-emissions for rapeseed, total GHGE from milk would increase by an average of about 2%—or even more in the case of LUC—for the conventional PS.

**Reducing energy required for mineral fertilizers and fuels.** Besides the production of renewable energy from biogas (as stated in ‘Manure management’ section), the application of fertilizers and related management measures deserve specific attention concerning their contribution to GHGE. Because emissions from soil increase when the available amount of N in the soil increases<sup>17</sup>, the quantity of nitrogen applied must be thoroughly adjusted to the requirements of plants. Furthermore, mineral fertilizers that need large amounts of (fossil) energy during production and transport should be substituted as

much as possible by livestock manure. On a long-term basis, housing and manure management systems should therefore be designed in a way that the emissions of nitrogen are kept as low as possible. According to IPCC<sup>17</sup> and calculations herein, pasture can be expected to protect nitrogen very well against emission processes. The use of fossil energy can also be reduced if feeding is generally based on the utilization of pastures and the avoidance of feedstuffs transported over long distances. As an example, the transport of soybeans from Brazil and of extracted soybean meal to be used in PS L<sub>con</sub> requires 12.3 liters of diesel per cow and year (0.57% of total GHGE).

## Conclusions

From the results presented herein and from information provided in the literature, it is concluded that organic milk PS are superior over conventional systems in terms of GHGE both per ha of farmland and per kg of milk. A relevant factor for these differences is LUC as a source for emissions, especially associated with soybean production in South America. For the systems considered in this study, the difference in GHGE per kg of milk between conventional and organic systems depends on the site-specific conditions for agricultural production: the higher the potential milk output per cow, the lower the differences that can be expected.

Regardless of the actual production system, the greatest proportion of GHGE originates from enteric fermentation. Although this inevitable source of emissions can be influenced quantitatively, clear limits exist for the degree of reduction that can be reached. Apart from enteric fermentation, manure management and forage supply also contribute substantially to GHGE. Consequently, dairy PS in which the focus lies on optimum forage quality, a high proportion of pasturage and additional fermentation of the manure in biogas plants will produce relatively low GHGE.

Despite the focus of this paper, the question of sustainable food production should not be restricted to factors that are currently discussed in connection with climate change. However, organic and low-input PS undoubtedly possess a number of strong points with regard to ecological and ethical aspects of sustainability.

## References

- 1 Gugele, B., Ritter, M., and Mareckova, K. 2002. Greenhouse gas emission trends in Europe, 1990–2000. Topic Report 7/2002. European Environment Agency, Copenhagen.
- 2 UBA (Umweltbundesamt; Austrian Environment Agency). 2004. Siebenter Umweltkontrollbericht – 5.5 Landwirtschaftlich genutzte Lebensräume. Available at Web site [http://www.umweltbundesamt.at/fileadmin/site/umweltkontrolle/2004/0505\\_land.pdf](http://www.umweltbundesamt.at/fileadmin/site/umweltkontrolle/2004/0505_land.pdf) (verified 23 November 2008; in German).
- 3 Luick, R. 1998. Ecological and socio-economic implications of livestock-keeping systems on extensive grasslands in

- south-western Germany. *Journal of Applied Ecology* 35:979–982.
- 4 Fehrenbach, H., Giegrich, J., Reinhardt, G., Schmitz, J., Sayer, U., Gretz, M., Seizinger, E., and Lanje, K. 2008. Criteria for a Sustainable Use of Bioenergy on a Global Scale. Report on behalf of the German Federal Environment Agency. Available at Web site [http://www.biofuelstp.eu/downloads/Criteria\\_for\\_sustainable\\_bioenergy\\_German\\_Research.pdf](http://www.biofuelstp.eu/downloads/Criteria_for_sustainable_bioenergy_German_Research.pdf) (verified 25 September 2008).
  - 5 IPCC (Intergovernmental Panel on Climate Change). 2007. *Climate Change 2007: Mitigation of Climate Change*. (Working Groups I, II, III). Available at Web site <http://www.ipcc.ch> (verified 11 June 2007).
  - 6 Garnett, T. 2009. Livestock-related greenhouse gas emissions: impacts and options for policy makers. *Environmental Science and Policy* 12:491–503.
  - 7 Löthe, K., Fuchs, F., and Zeddies, J. 1997. Reduction of emissions in farming systems in Germany. In W.N. Adger, D. Pettenella, and M. Whitby (eds). *Climate-change Mitigation and European Land-use Policies*. CAB International, London, UK. p. 159–169.
  - 8 Thomassen, M.A., van Calker, K.J., Smits, M.C.J., Iepema, G.L., and de Boer, I.J.M. 2008. Life cycle assessment of conventional and organic milk production in the Netherlands. *Agricultural Systems* 96:95–107.
  - 9 Williams, A.G., Audsley, E., and Sandars, D.L. 2006. Determining the environmental burdens and resource use in the production of agricultural and horticultural commodities. In Main Report, Defra Research Project IS0205, Cranfield University and Defra. Available from Web site <http://www.defra.gov.uk> (verified 25 September 2008).
  - 10 Cederberg, C. and Mattson, B. 2000. Life cycle assessment of milk production – a comparison of conventional and organic farming. *Journal of Cleaner Production* 8:49–60.
  - 11 IPCC (Intergovernmental Panel on Climate Change). 2001. *Climate Change 2001: The Scientific Background*. Cambridge University Press, Cambridge, UK.
  - 12 Invekos (Integriertes Verwaltungs- und Kontrollsystem; IACS, Integrated Administration and Control System). 2009. Database, Austrian Federal Ministry of Agriculture, Forestry, Environment and Water Management, Vienna.
  - 13 Kirner, L. 2009. Vollweide in der Bio-Milchviehhaltung aus ökonomischer Sicht am Beispiel Österreichs. In Proceedings of 10. Wissenschaftstagung Ökologischer Landbau, Part 2. ETH-Zürich, Zürich, p. 250–253 (in German).
  - 14 BMLFUW (Bundesministerium für Land- und Forstwirtschaft, Umwelt und Wasserwirtschaft; Austrian Federal Ministry of Agriculture, Forestry, Environment and Water Management). 2006. *Grüner Bericht 2006*. Available from Web site <http://www.gruenerbericht.at> (verified 2 April 2007; in German).
  - 15 Buchgraber, K. and Gindl, G. 2004. *Zeitgemäße Grünlandbewirtschaftung*. 2. Auflage. Leopold Stocker Verlag, Graz, Austria (in German).
  - 16 BMLFUW (Bundesministerium für Land- und Forstwirtschaft, Umwelt und Wasserwirtschaft; Austrian Federal Ministry of Agriculture, Forestry, Environment and Water Management). 2008. *Deckungsbeiträge und Daten für die Betriebsplanung 2008*. BMLFUW, Vienna, Austria (in German).
  - 17 IPCC (Intergovernmental Panel on Climate Change). 2006. In H.S. Eggleston, L. Buendia, K. Miwa, T. Ngara, and K. Tanabe (eds). *Guidelines for National Greenhouse Gas Inventories*. Prepared by the National Greenhouse Gas Inventories Programme. IGES, Japan.
  - 18 Rahmann, G. and Böhm, H. 2005. Organic fodder production in intensive organic livestock production in Europe: recent scientific findings and the impact on the development of organic farming. In Rowlinson et al. (eds). *Integrating Livestock–Crop Systems to Meet the Challenges of Globalisation*. Proceedings of the AHAT/BSAS International Conference, 14–18 November 2005 in Khon Kaen, Thailand. Volume 1. British Society of Animal Science, Penicuik, UK. p. 471–485.
  - 19 DLG (Deutsche Landwirtschafts-Gesellschaft). 1997. *DLG-Futterwerttabelle für Wiederkäuer*. 7. erweiterte und überarbeitete Auflage. DLG-Verlag, Frankfurt/Main, Germany (in German).
  - 20 Buchgraber, K., Gruber, L., Resch, R., and Wiedner, G. 1997. *ÖAG-Futterwerttabelle für Grundfutter im Alpenraum*. In *ÖAG-Sonderbeilage – Der Fortschrittliche Landwirt*, 4/1997 (in German).
  - 21 Resch, R., Guggenberger, T., Wiedner, G., Kasal, A., Wurm, K., Gruber, L., Ringdorfer, F., and Buchgraber, K. 2006. *Futterwerttabellen im Jahr 2006 für das Grundfutter im Alpenraum*. Available from Web site <http://www.gumpenstein.at> (verified 22 May 2008; in German).
  - 22 Amon, B., Fröhlich, M., Weißensteiner, R., Zablatnik, B., and Amon, T. 2007. *Tierhaltung und Wirtschaftsdüngermanagement in Österreich (TIHALO)*. Final Report Project No. 1441 on behalf of the Federal Ministry of Agriculture, Forestry, Environment and Water Management, Vienna, Austria (in German).
  - 23 Amon, B., Fröhlich, M., Weißensteiner, R., Zablatnik, B., and Amon, T. 2007. Database ‘TIHALO’, unpublished data.
  - 24 Van Es, A.J.H. 1975. Feed evaluation for dairy cows. *Livestock Production Science* 2:95–101.
  - 25 GfE (Gesellschaft für Ernährungsphysiologie). 2001. *Empfehlungen zu Energie- und Nährstoffversorgung der Milchkühe und Aufzuchttrinder*. DLG-Verlag, Frankfurt/Main, Germany (in German).
  - 26 Statistics Austria. 2005. In BMGF (Bundesministerium für Gesundheit und Frauen), 2006: *Herkunftsidentität von Raps und Rapsprodukten am Markt in Österreich und Verarbeitung in dezentralen Ölmöhlen*. Vienna, Austria (in German).
  - 27 AGES (Agentur für Gesundheit und Ernährungssicherheit; Austrian Agency for Health and Food Safety). 2005. *Machbarkeitsstudie für Auslobung “gentechnikfrei” und Vermeidung von GVO bei Lebensmittel aus tierischer Erzeugung*. Vienna, Austria (in German).
  - 28 Zuchtdata (Association of Austrian Cattle Breeders). 2007. *Annual Report 2006*. Available from Web site <http://www.zar.at/download/RiZu/ZuchtData-Jahresbericht-2006.pdf> (verified 15 July 2007; in German).
  - 29 Kirchgeßner, M., Windisch, W., and Müller, H.L. 1995. Nutritional factors for the quantification of methane production. In W.V. Engelhardt, S. Leonhard-Marek, G. Breves, and D. Giesecke (eds). *Ruminant Physiology: Digestion, Metabolism, Growth and Reproduction*. Proceedings of the VIII International Symposium on Ruminant Physiology. p. 333–348.
  - 30 Öhlinger, C., Dell, G., and Egger, C. 2007. *Stromsparen in der Landwirtschaft*. Manual, Energy Agency of Upper Austria. Available from Web site [http://www.wsed.at/wsed/fileadmin/esv\\_files/Info\\_und\\_Service/Landwirtschaftsbro-21.12.07-fin.pdf](http://www.wsed.at/wsed/fileadmin/esv_files/Info_und_Service/Landwirtschaftsbro-21.12.07-fin.pdf) (verified 20 January 2008; in German).

- 31 Ecoinvent. 2007. Ecoinvent Data v2.0 Final Reports Ecoinvent 2000, No. 1–25, Swiss Centre for Life Cycle Inventories, Dübendorf.
- 32 ÖKL (Österreichisches Kuratorium für Landtechnik und Landentwicklung; Austrian Council for Agricultural Engineering and Rural Development). 2005. ÖKL- Richtwerte für die Maschinenselbstkosten 2005. ÖKL, Vienna, Austria (in German).
- 33 Wiltling, H.C., Benders, R.M.J., Kok, R., Biesiot, W., and Moll, H.C. 2004. EAP – Energy Analysis Program. 2nd revised ed. IVEM Research Report 98, Groningen, The Netherlands.
- 34 Patyk, A. and Reinhardt, G.A. 1997. Düngemittel-, Energie und Stoffstrombilanzen. Vieweg-Verlag, Braunschweig/Wiesbaden, Germany (in German).
- 35 Biskupek, B., Patyk, A., and Radtke, J. 1997. Daten zu Pflanzenproduktion. In M. Kaltschmitt and G.A. Reinhardt (eds). *Nachwachsende Energieträger*. Vieweg-Verlag, Braunschweig/Wiesbaden, Germany (in German).
- 36 Gruber, L. and Steinwider, A. 1996. Einfluß der Fütterung auf die Stickstoff- und Phosphorauscheidung landwirtschaftlicher Nutztiere – Modellkalkulationen auf Basis einer Literaturübersicht. In *Die Bodenkultur – Austrian Journal of Agricultural Research*, 47(4). WUV-Universitätsverlag, Vienna, Austria (in German).
- 37 Schechtner, G. 1991. Wirtschaftsdünger – Richtige Gewinnung und Anwendung, Sonderausgabe des Förderungsdienst 1991. Bundesministerium für Land- und Forstwirtschaft, Vienna, Austria (in German).
- 38 Pötsch, E.M. 2006. Österreichisches Aktionsprogramm zur Umsetzung der EU-Nitratrichtlinie: Aktualisierung der N-Ausscheidungsrate für landwirtschaftliche Nutztiere – Konsequenzen für die Praxis. Available from Web site [http://www.gumpenstein.at/publikationen/umweltprogramme\\_2006/poetsch.pdf](http://www.gumpenstein.at/publikationen/umweltprogramme_2006/poetsch.pdf) (verified 21 October 2007; in German).
- 39 EEA (European Environment Agency). 2008. Land cover changes 1990–2000. Available from Web site <http://data-service.eea.europa.eu/dataservice/viewdata/viewpvt.asp> (verified 8 October 2008).
- 40 Küstermann, B., Kainz, M., and Hülsbergen, K.-J. 2007. Modelling carbon cycles and estimation of greenhouse gas emissions from organic and conventional farming systems. *Renewable Agriculture and Food Systems* 23:1–16.
- 41 Fliessbach, A., Oberholzer, H.-R., Gunst, L., and Mäder, P. 2007. Soil organic matter and biological soil quality indicators after 21 years of organic and conventional farming. *Agriculture Ecosystems and Environment* 118:273–284.
- 42 Soussana, J.-F., Loiseau, P., Vuichard, N., Ceschia, E., Balesdent, J., Chevallier, T., and Arrouays, D. 2004. Carbon cycling and sequestration opportunities in temperate grasslands. *Soil Use and Management* 20:219–230.
- 43 Kirchgeßner, M. 2004. Tierernährung. 11. Überarbeitete Auflage. DLG- Verlag, Frankfurt/Main, Germany (in German).
- 44 Fürst, C. 2006. Züchterische Strategien für die Bio-Rinderzucht. In *Bericht über die österreichische Fachtagung für Biologische Landwirtschaft 2006*. Available from Web site <http://orgprints.org/8363/01/Tagungsband.pdf> (verified 15 December 2007; in German).
- 45 Hindrichsen, I.K., Wettstein, H.-R., Machmüller, A., Jörg, B., and Kreuzer, M. 2005. Effect of the carbohydrate composition of feed concentrates on methane emission from dairy cows and their slurry. *Environmental Monitoring and Assessment* 107:329–350.
- 46 Lehuger, S., Gabrielle, B., and Gagnaire, N. 2009. Environmental impact of the substitution of imported soybean meal with locally produced rapeseed meal in dairy cow feed. *Journal of Cleaner Production* 17:616–624.
- 47 Dalgaard, R., Schmidt, J., Halberg, N., Christensen, P., Thrane, M., and Pengue W.A. 2008. LCA of Soybean Meal. *International Journal of LCA* 13(3):240–254.
- 48 Renewable Fuels Agency (Department for Transport). 2008. Carbon and sustainability reporting with the Renewable Transport Fuels Obligation – Technical Guidance (Part 2). Renewable Fuels Agency (Department for Transport), London, UK.
- 49 Smaling, E.M.A., Roscoe, R., Lesschen, J.P., Bouwman, A.F., and Comunello, E. 2008. From forest to waste: Assessment of the Brazilian soybean chain, using nitrogen as marker. *Agriculture, Ecosystems and Environment* 128:185–197.
- 50 Weiske, A., Vabitsch, A., Olesen, J.E., Schelde, K., Michel, J., Friedrich, R., and Kaltschmitt, M. 2006. Mitigation of greenhouse gas emissions in European conventional and organic dairy farming. *Agriculture, Ecosystems and Environment* 112:221–232.
- 51 Olesen, J.E., Schelde, K., Weiske, A., Weisbjerg, M.R., Asberg, W.A.H., and Djurhuus, J. 2006. Modelling greenhouse gas emissions from European conventional and organic dairy farms. *Agriculture, Ecosystems and Environment* 112:207–220.
- 52 IPCC (Intergovernmental Panel on Climate Change). 1997. Revised 1996 IPCC Guidelines for National Greenhouse Gas Inventories. OECD, Paris.
- 53 Fleischer, P., Metzner, M., Beyerbach, M., Hoedemaker, M., and Klee, W. 2001. The relationship between milk yield and the incidence of some diseases in dairy cows. *Journal of Dairy Science* 84:2025–2035.
- 54 Van Soest, P. 1994. *Nutritional Ecology of the Ruminant*. Cornell University Press, Ithaca, NY, USA.
- 55 Weyermann, I., Kampmann, D., Peter, M., Herzog, F., and Lüscher, A. 2006. Mountain meadows have a high ecological quality. *Agrarforschung* 13(04):156–161.
- 56 Kampmann, D., Herzog, F., Jeanneret, Ph., Konold, W., Peter, M., Walter, T., Wildi, O., and Lüscher, A. 2008. Mountain grassland biodiversity: impact of site conditions versus management type. *Journal for Nature Conservation* 16:12–25.
- 57 Knaus, W. 2008. Dairy cows trapped between performance demands and adaptability. *Journal of the Science of Food and Agriculture* 89:1107–1114.
- 58 Tuytens, F.A.M. 2005. The importance of straw for pig and cattle welfare: a review. *Applied Animal Behaviour Science* 92:261–282.
- 59 Amon, B. 2006. Emissionen umwelt- und klimarelevanter Gase aus der landwirtschaftlichen Tierhaltung. Habilitation thesis, University of Natural Resources and Applied Life Sciences, Vienna, Austria (in German).
- 60 KTBL (Kuratorium für Technik und Bauwesen in der Landwirtschaft; German Association for Technology and Structures in Agriculture). 2005. *Gasausbeute in Landwirtschaftlichen Biogasanlagen*. KTBL-Heft 50. Darmstadt, Germany (in German).
- 61 Amon, B., Kryvoruchko, V., Amon, T., and Zechmeister-Boltenstern, S. 2006. Methane, nitrous oxide and ammonia emissions during storage and after application of dairy cattle

- slurry and influence of slurry treatment. *Agriculture, Ecosystems and Environment* 112:153–162.
- 62 UN-Energy. 2007. Sustainable Bioenergy. A Framework for Decision Makers. Available from Web site <http://esa.un.org/un-energy/pdf/susdev.Biofuels.FAO.pdf> (verified 15 March 2009).
- 63 The Royal Society. 2008. Sustainable Biofuels: Prospects and Challenges. The Clyvedon Press. Available from Web site <http://royalsociety.org/displaypagedoc.asp?id=28914> (verified 25 January 2008).
- 64 Crutzen, P.J., Mosier, A.R., Smith, K.A., and Winiwarter, W. 2007. N<sub>2</sub>O release from agro-biofuel production negates global warming reduction by replacing fossil fuels. Available from Web site <http://www.atmos-chem-phys-discuss.net/7/11191/2007/acpd-7-11191-2007.pdf> (verified 21 January 2008).